

The Distribution Behavior of Stable and Radioactive Cs between the Clastic Grains and Water in the Isolated Undisturbed Mountain Pond

Kazumi IWAMOTO¹, Makio KIMURA¹, Kennji ANDO¹, Hikaru AMANO²,
Hirotake MORIYAMA³, Keizo KAWAMOTO⁴, and Jituya TAKADA⁴

The concentrations of stable Cs and radioactive fallout ^{137}Cs were determined for the clastic grain of the isolated undisturbed pond "Yasha-ga-Ike", located at an elevation of 1100m near the ridge of mountain in Fukui Pref. The concentrations of both cesiums varied with particle size and increased with a decrease in diameter. Using the concentration data for the pond water, the in-situ distribution coefficient K_d defined by the ratio of the element concentration adsorbed on the particle to that of the water was obtained both for the stable Cs of eight samples, taken at different times from June 1999 to October 2001, and the radioactive ^{137}Cs of four samples. The values of K_d increased with a decrease in particle diameter and ranged from about 10 to 1000m³/kg, and the K_d for the stable Cs were slightly larger than those of ^{137}Cs . The results suggest that the adsorption of the stable Cs was established before the fallout of the ^{137}Cs , and the isotopic exchange between both cesiums has been proceeding from the water to the adsorbed layer and the grain matrix.

1. Introduction

The distribution coefficient, which is defined as partitioning of a solute between a solid and liquid phase, is used in various models which describe the migration and transport phenomena of radionuclides in environment and is one of the important parameters for environment safety assessment of nuclear waste disposal. The data of the coefficients are obtained mostly through in-vitro batch or column experiment using radioactive tracer and less through in-situ measurement in the environment⁽¹⁻⁵⁾. As to the behavior of caesium in the environment, the migration and sorption in the lake or river have been studied only on the radioactive caesium derived mostly from the weapon-testings and Chernobyl accident⁽⁶⁻¹⁰⁾, and the studies on both radioactive and stable caesiums are found in the transfer and accumulation in such plant ecosystem as the soil-plant and/or fungus-animal pathway⁽¹¹⁻¹²⁾. Thus the studies for obtaining the distribution coefficients of both caesiums from a natural appropriate environment are few. This study concerns the simultaneous measurement of in-situ distribution coefficients of both radioactive ^{137}Cs and stable Cs between clastic grains and pond water in the small isolated undisturbed pond named "Yasha-ga-Ike", where the partitioning equilibrium in the pond was regarded to be attained in a long period.

The researches on the water and the grains of this pond have been carried out as in the following. The samples of grains and pond water were taken near the beach and the concentrations of elements were analyzed by the neutron activation and the ICP-MS method, respectively. The concentration of any constituent element of the grain was found to decrease with increase in particle size, and the adsorptive behavior of trace elements onto the grains was postulated⁽¹³⁾ and then confirmed⁽¹⁴⁾. The concentrations of radioactive ^{137}Cs from the global fallout of the weapon testings in the sixties and stable Cs in the pond water were determined, as will be referred later⁽¹⁵⁾. It is the purpose of the present study to elucidate the distribution behavior of radioactive ^{137}Cs and stable Cs between the clastic grains and water, and the in-situ distribution coefficients of both

1 Department of Environmental and Biotechnological Frontier Engineering, Fukui Univ. of Technology

2 Japan Atomic Energy Research Inst. 3 Faculty of Eng., Kyoto Univ. 4 Research Reactor Inst., Kyoto Univ.

cesiums between the grains and the pond water were obtained and compared.

2. Experimental Procedures

2.1. Small Isolated Undisturbed Pond "Yasha-ga-Ike"

The pond "Yasha-ga-Ike", with maximum depth of 7.5m and about 40are of water surface is located at an elevation of 1100m near the ridge of mountain in Fukui Pref. The pond is supposed to be formed by landslide of the surrounding mountain of chert formation in the latter period of the diluvial epoch. Three sides of the pond is surrounded by steep slopes except one side covered with bush. The watershed of the pond is small and none of rivers flow in and out. The pond water is usually clear having the turbidity of about 2mg/l at deepest point except the rainy day and is said to have never run dry. A small Japanese diving beetle "Yasha-Gengoro" specially protected due to its rareness lives in the water. Thus this pond is regarded as an isolated undisturbed non-volcanic pond, and it is supposed that all the fallouts from the sky have been accumulated. The outline of the pond is shown in Fig. 1, in which a bold line indicates the contour of water surface, a thin line the beach, a dotted line the portion of mud sediments and a bold double dotted line the ridge line deviding Fukui Pref. and Gifu Pref. As to the sediments of the pond, mud covers the central deep location, and clastic grains of chert with different sizes are found both at the shallow place in the pond and the flat beach which is covered with water and snow in winter time.

As described in the following, the samples of amber-colored clastic grains and pond water were taken, and the concentrations of stable Cs and radioactive ^{137}Cs were determined by the neutron activation, the ICP-MS or the gamma spectrometry. Both the grain surface rubbing-off and the nitric acid leach techniques were applied to the grain samples to confirm the adsorptive behavior of both cesiums. After determining the adsorbed amounts of them, the distribution coefficients between the grains and the pond water were obtained.

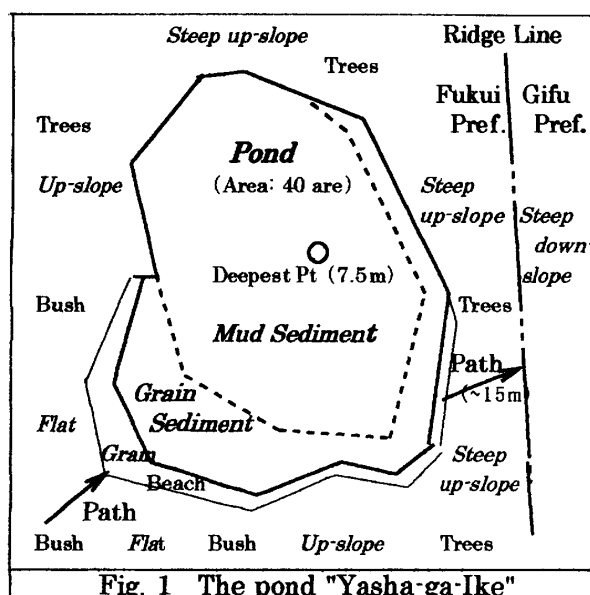


Fig. 1 The pond "Yasha-ga-Ike"

2.2. Measurements of Stable and Radioactive Cs Concentrations for the Grain and Water Samples

Seven samples of grains and pond water from the shallow place of about 50cm depth near the beach and one grain sample from the beach were taken at different times from June 1999 to October 2001 as given in Table 1, where the temperature of water varied from 12.5 to 25.5 °C. The grain samples were dried and sorted into 11 fractions of diameters: 2-1mm, 1mm-710µm, 710-500, 500-355, 355-250, 250-180, 180-125, 125-90, 90-63, 63-38 and < 38µm. The sieves were chosen so as to meet the ratio of the opening dimension of any screen to that of the next smaller screen is square root of two. For each fractions of sorted grain samples, neutron activation analysis and ^{137}Cs radioactivity measurement were carried out. The samples of about 80 mg were irradiated for 20 sec and 60 min under thermal neutron flux of 2.3×10^{23} and 2.8×10^{13} n/cm²·s, respectively, by KURRI (Kyoto University Research Reactor) and the gamma spectrometry was applied in

order to obtain the concentrations of about 25 elements such as Na, K, Rb, Cs, Mg, Ba, Cr, Mn, Fe, Co, Al, Sc, La, Ce etc., although only the Cs is referred in this study. The gamma-ray spectrometry by a Ge(Li) detector was applied to each particle sample of about 60g contained in the polystyrene vessel with 6cm diameter by 2cm high, and the concentration of ^{137}Cs was measured. The decay corrections for the samples were made as to February 2000, when the measurement of DB sample was done.

For the water samples, the concentrations of stable Cs was measured by the ICP-MS, and the radioactive ^{137}Cs collected in-situ at the pond by pumping the water to the packed column of acrylic fibre adsorbent, impregnated with copper hexacyanocobalt ferrate⁽⁶⁾, was analyzed by the gamma-ray spectrometry, as described in the previous report⁽⁵⁾.

2.3. The Grain Surface Rubbing-off and the Nitric Acid Leach of the Grain Samples

When the cesium element is adsorbed on the grain surface, the concentration of cesium on and near the surface is expected to be higher than that of the grain matrix. To ascertain this expectation directly, the following two methods of the surface rubbing-off and the nitric acid leach are applied to the sieved particles. In case of the surface rubbing-off technique, the particles of HB1 and FB1 samples with size range of 2-1 mm in diameter are rubbed gently by using agate mortar and pestle and then sieved. The concentrations of stable Cs and ^{137}Cs in the sieved particles are analyzed by the neutron activation and the gamma spectrometry, respectively, and compared. In case of the nitric acid leach, the sieved particle samples of IB and LB that have finished the measurement of ^{137}Cs radioactivity are leached with 8N nitric acid near the boiling point for 3 hours. After the leach, the concentrations of stable Cs and ^{137}Cs in the residual particles were measured and the concentration leached was obtained as the difference between the concentrations before and after the leach. Also the concentration of stable Cs in the acid solution after the leach is measured by the ICP-MS, where the solution was diluted by 1000 times, to obtain directly the concentration leached.

3. Experimental Results

3.1. Concentrations of Stable Cs and Radioactive ^{137}Cs in the Grains

Table 1 The Grains and Water Samples and Notations

Exp No.	Sampling Date	Water Temp. (°C)	For Stable Cs		For ^{137}Cs	
			Grain	Water	Grain	Water
D	June 9, '99	18.0	DB	DW	DB*	DW*
F	Oct. 16, '99	15.0	FB	FW	FB*	
G	Nov. 07, '99	12.5	GB	GW	GB*	GW*
H	May 23, '00	19.0	HB	HW	HB*	
I	July 17, '00	25.0	IB, IY	IW	IB*	IW*
J	Oct. 14, '00	15.5	JB	JW		
K	May 20, '01	20.0	KB	KW		
L	July 23, '01	25.5	LB	LW	LB*	LW*
N	Oct. 13, '01	14.0	NB	NW	NB*	

Note : IY sample was taken from the beach.

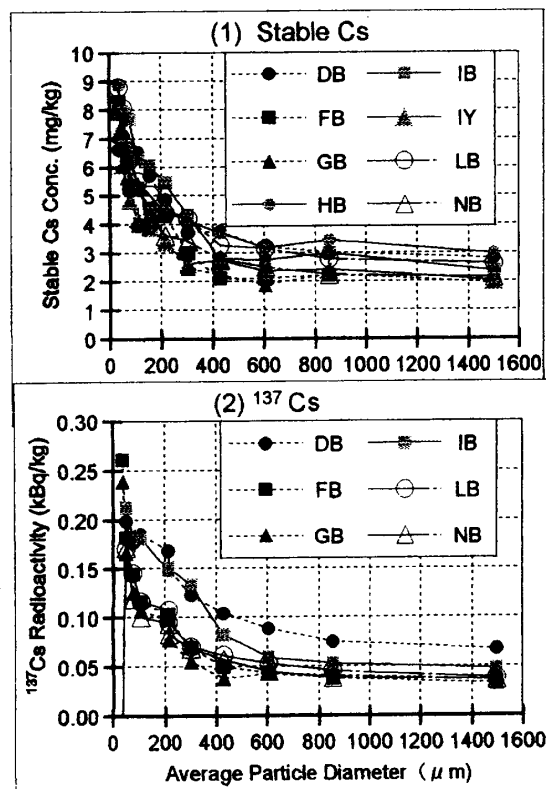


Fig 2 The concentrations of Stable Cs and ^{137}Cs in the Grain Samples

The concentrations of stable Cs for the eight grain samples of DB, FB, GB, HB, IB, IY, LB and NB are plotted against the average diameter of the sieved particles in Fig. 2(1) and those of ^{137}Cs for the six samples of DB*, FB*, GB*, IB*, LB* and NB* in Fig. 2(2). It is seen that the concentrations of both stable Cs and ^{137}Cs decrease with increase in diameter, as reported for many stable elements in cases of DB and GB samples in the previous report¹⁶⁾, and that the difference of concentrations among samples is not so large.

The concentration of stable Cs or ^{137}Cs in the grain matrix was obtained by extrapolating the grain size to an infinitive. The concentration of stable Cs is 2.4mg/kg for DB, 1.5 for FB, 1.3 for GB, 2.4 for HB, 2.0 for IB, 1.5 for IY, 2.0 for LB and 1.5 for NB, and that of ^{137}Cs is 0.080kBq/kg for DB*, 0.030 for FB*, 0.025 for GB*, 0.035 for IB*, 0.025 LB* and 0.025 for NB*, respectively.

3.2. The Grain Surface Rubbing-off and the Nitric Acid Leach of the Grain Samples

(1) The Grain Surface Rubbing-off

After the surface rubbing-off of the particles of HB1 and FB1 samples with 2-1 mm in diameter, the concentrations of stable Cs and ^{137}Cs were measured for the particles size ranges of 2-1mm, 710-500, 355-250, 180-125 and 90-63 μm . The results are shown in Fig. 3, where the ratio of the concentration of the residual particles after the rubbing to that of the original HB1 or FB1 particles, denoted as X:1500, is plotted against the diameter in order to make the effect of the rubbing clear. It is seen that the concentration of the residual particle, whose surface was rubbed-off and shown by Y:1500, are lower than those of the fine particles which are derived mostly from the near-surface and shown as Y:605, Y:303, Y:153 and Y:77 corresponding to the size range. Thus the concentration on and near the particle surface was confirmed to be higher than that of the inner grain matrix.

(2) The Nitric Acid Leach

After the nitric acid leach, the concentrations of stable Cs both in the residual particles and the leached solution and those of radioactive ^{137}Cs in the residual particles were measured, and the results for the sample LB are shown in Fig. 4. In this figure, such concentrations as the whole cesium in the particle before the leach, the adsorbed cesium, the leached cesium and the residual cesium after the leach are plotted against the average particle diameter. The concentration of the adsorbed cesium was calculated by subtracting the concentration of grain matrix from the whole concentration. The concentration of the leached cesium was determined by the analysis of the leached solution by the ICP-MS and that of ^{137}Cs by taking the difference of gamma radioactivities of the particles before and after the leach. The similar results were obtained for the sample IB.

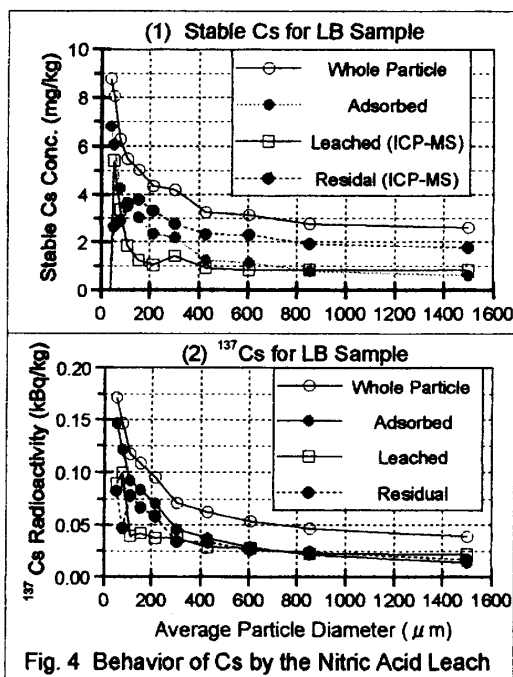
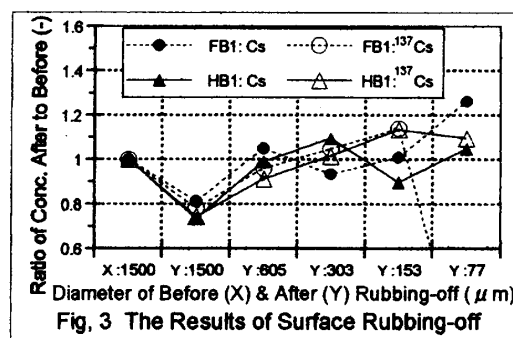


Fig. 4 Behavior of Cs by the Nitric Acid Leach

The followings results were obtained from the nitric acid leach: (1) The leached concentrations of stable Cs and ^{137}Cs decreased with increase in particle diameter and they are fairly close to the adsorbed concentrations over a wide range of the particle diameter.

(2) The residual concentrations after the leach show fairly uniform values of about 2.1mg/kg for stable Cs and 0.025kBq/kg for ^{137}Cs for the sample IB and 2.0mg/kg and 0.025kBq/kg for the sample LB, respectively, except the values for smaller diameters. These values are pretty close to the particle matrix concentrations, which are obtained as 2.0mg/kg for stable Cs and 0.035kBq/kg for ^{137}Cs for the sample IB and 2.0mg/kg and 0.025kBq/kg for the sample LB, respectively.

4. Discussions

4.1. Adsorptive Behavior of Cs onto the Grains

In the previous study¹⁴⁾, both techniques of the surface rubbing-off and the nitric acid leach described before were applied to the grain samples and the adsorptive behavior of non-radioactive trace elements was confirmed. The experimental results obtained in this study showed that (1) the concentrations of both stable Cs and ^{137}Cs on and near the particle surface are clearly higher than that of the particle matrix and (2) the concentrations of leached stable Cs or ^{137}Cs are fairly close to that of each cesium which was supposed to have adsorbed over a wide range of particle diameter. These results are considered to give more direct evidences of adsorption on the surface of particles rather than such mechanism as the diffusion into the grain. Furthermore, the fact that the nitric acid leach was necessary to strip off the adsorbed Cs from the grain indicates that the adsorbed Cs on the grain are so tightly adhered to the surface. It is to be pointed out here that the chemical adsorptive behavior of the elements on the grain is still remained to be elucidated in future.

Since the adsorptive behavior is thus confirmed, the concentration of the adsorbed cesium can be obtained by subtracting the concentration of grain matrix, which is estimated by extrapolating the particle size to an infinitive, from the whole concentration. The matrix concentration is a constant among samples in principle, but the concentrations estimated varied from 1.3 to 2.4 mg/kg for the stable Cs and 0.025 to 0.060 kBq/kg for the ^{137}Cs as described before, and the average values are about 2.0 mg/kg for the stable Cs and about 0.03 kBq/kg for ^{137}Cs . All the data of the adsorbed concentrations are shown in Fig. 6 for stable Cs of eight samples and for ^{137}Cs of six samples. It is to be noted in this figure that the adsorbed concentrations of both stable Cs and ^{137}Cs are pretty close among samples although they were taken under different conditions of season and temperature as given in Table 1. The adsorbed concentrations for the IY sample seem to be slightly smaller than those of IB sample. This

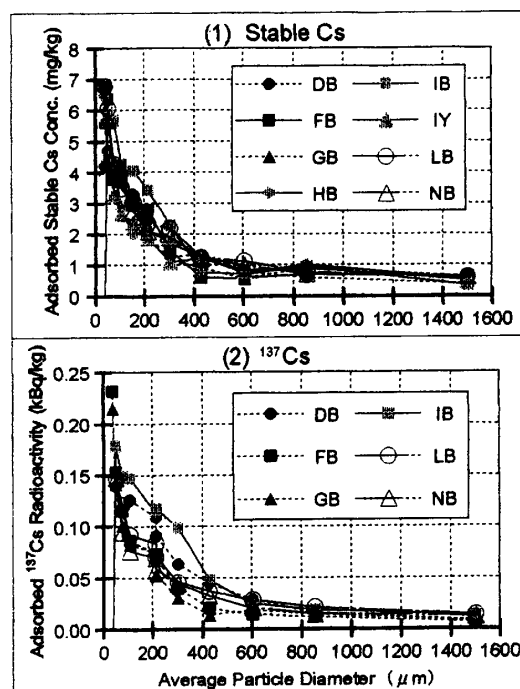


Fig. 5 The Concentrations of Adsorbed Stable Cs and ^{137}Cs in the Grain Samples

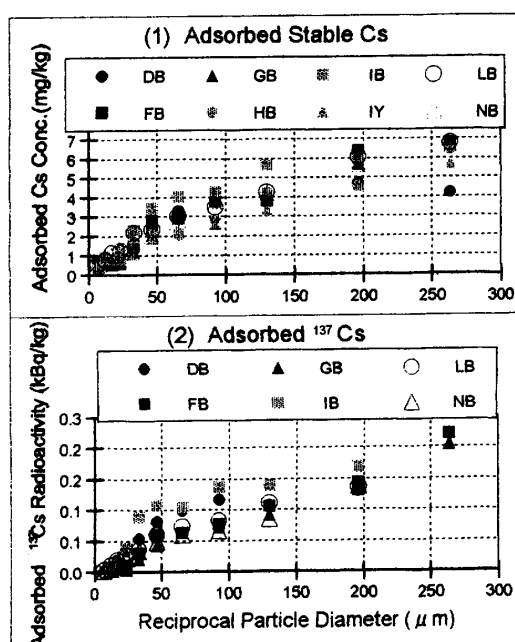


Fig. 6 Plots of Cs Conc. against Reciprocal of Particle Diameter

may come from the fact that the IY sample was taken from the beach where is covered with water only in the winter time.

Furthermore, when the concentrations of stable Cs and ^{137}Cs are plotted against the reciprocal of the particle diameter, the linear relationship that intercepts the zero point of the co-ordinate axis is not found in case of the whole cesium concentrations, but found in case of the adsorbed cesium concentrations as shown in Fig. 6. This result suggests that the migration of cesium from the water to the grains is based on the adsorption rather than the mere diffusion into the grain matrix, and the concentration of adsorbed cesium is roughly proportional to the surface area of the particles.

4.2. In-situ Distribution Coefficients of Stable Cs and ^{137}Cs

Since the concentrations of stable Cs and ^{137}Cs adsorbed on the particles are obtained and their concentrations in the pond water were already reported as quoted in Table 2¹⁵⁾, the distribution coefficient K_d defined by the ratio of the element concentration adsorbed on the particle to that of pond water is obtainable. The in-situ distribution coefficients of

Table 2 Concentrations of Stable Cs and Radioactive ^{137}Cs in the Water Samples¹⁵⁾
(Concentration unit is mg/m^3 for stable Cs and Bq/m^3 for ^{137}Cs)

Sample	D	F	G	H	I	J	K	L	N
Stable Cs	0.0050	0.0089	0.0055	0.0122	0.0147	0.0108	0.0110	0.018	0.013
Rad. ^{137}Cs	0.23	-	0.27	-	0.85	-	0.84	-	-

stable Cs for the samples DB, FB, GB, HB, IB, IY, LB and NB and of radioactive ^{137}Cs for the samples DB, GB, IB and LB are plotted against the particle diameter in Fig. 7. It is seen for both cases of stable Cs and radioactive ^{137}Cs that (1) the K_d depends largely on particle diameter and increases with the decrease in diameter, (2) the change of K_d among samples is small, although the samples were taken at different conditions, (3) the K_d of ^{137}Cs seems to be slightly smaller than that of stable Cs and (4) the values of K_d ranged from about 20 to $1200\text{m}^3/\text{kg}$ for stable Cs and about 15 to $1000\text{m}^3/\text{kg}$ for ^{137}Cs . The K_d seems to depend largely on the concentration in the water, since the adsorbed concentrations do not differ largely among samples as seen in Fig. 5.

4.3. Comparison of Distribution Coefficients of Stable Cs and ^{137}Cs

The in-situ distribution coefficients of stable Cs and radioactive ^{137}Cs are compared in Fig. 8 for the samples of DB, GB, IB and LB, respectively. The in-situ K_d s for the radioactive ^{137}Cs (shown with *) seem to be slightly smaller than those for the stable Cs. In order to elucidate until what degree the radioactive ^{137}Cs migrated into the grain during these about forty years after the nuclear weapon testings in the atmosphere, it is interesting to compare the ratios of ^{137}Cs to stable Cs concentration in the whole particle, in the adsorbed layer and in the grain matrix. The variations of the ratios with particle

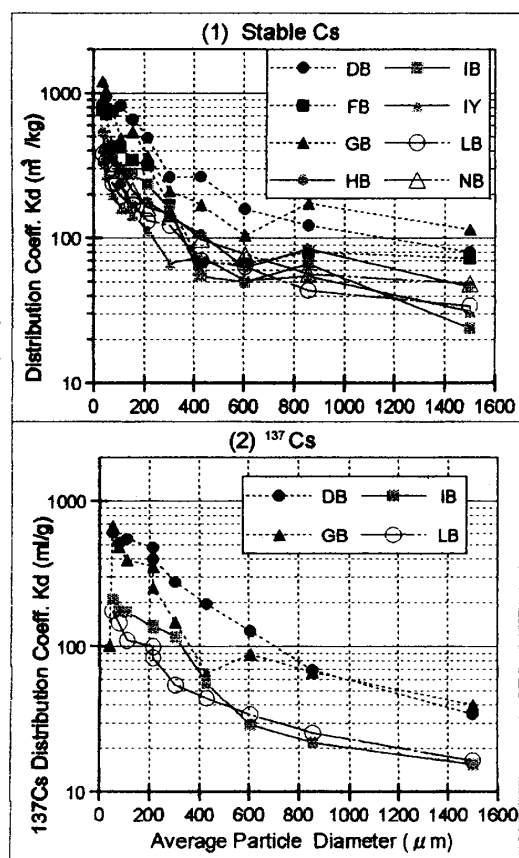


Fig. 7 In-situ Distribution Coefficient K_d of Stable Cs and ^{137}Cs

diameter are shown in Fig. 9 for the whole particle and the adsorbed layer. The ratios seem to decrease slightly with increase in diameter, and the degree of the decrease is larger in case of the adsorbed. The average values of the ratios, ignoring the effect of diameter dependence, for the whole particle, the adsorbed layer and the grain matrix are compared among the samples in Fig. 10. It is seen that the ratio of the adsorbed layer is larger than that of the matrix, and the average values of the ratios for all the samples are 23 for the whole particle, 26 for the adsorbed layer and 19 MBq/kg for the grain matrix. This result may be explained in the following way: the adsorption of stable Cs was held before the deposition of the ^{137}Cs and the isotopic exchange between ^{137}Cs and stable Cs has been proceeding from the water to the adsorbed layer preferably and to the matrix partially during these forty years. Furthermore, the ratio for the pond water was 47MBq/kg and is twice of the particle, which suggests that the isotopic exchange between the stable Cs and radioactive ^{137}Cs is still proceeding.

5. Conclusions

The concentrations of stable Cs and radioactive fallout ^{137}Cs were determined for the clastic grains of chert in the isolated undisturbed mountain pond "Yasha-ga-Ike", located at the elevation of 1100m near the ridge of mountain in Fukui Pref. The concentration of each cesium was found to decrease with increase in the particle size. After confirming their adsorptive behavior by surface rubbing-off and nitric acid leach techniques, their adsorbed concentrations on the grain were determined by subtracting the concentration of grain matrix. These concentrations increased with decrease in particle size and the variation among samples was small. The in-situ distribution coefficients K_d between the particles and the pond water were obtained for the stable Cs of eight samples and for ^{137}Cs of four samples. The followings were elucidated: (1) The values of K_d depend largely on particles diameter and increase with the decrease in diameter from about 10 to $100\text{m}^3/\text{kg}$, (2) the K_d depends less on the kind of the samples, (3) the K_d of ^{137}Cs seems to be slightly smaller than that of stable Cs. It was considered that the adsorption of stable Cs was held before the deposition of fallout ^{137}Cs .

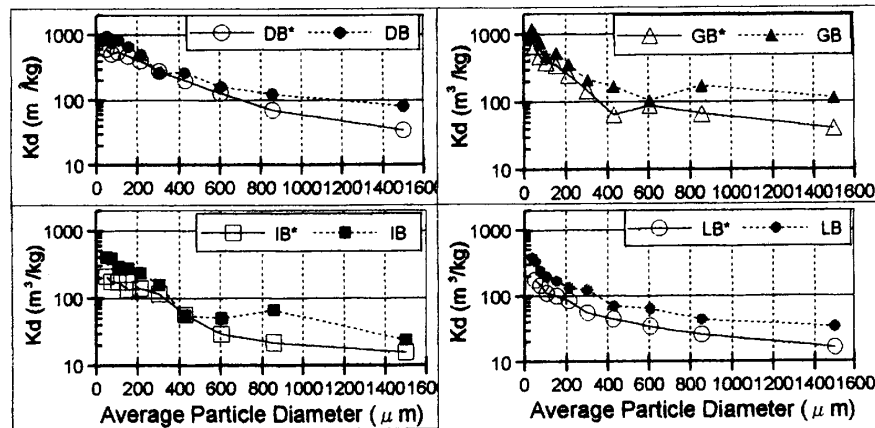


Fig. 8 Comparison of K_d between Stable Cs and ^{137}Cs (shown with*)

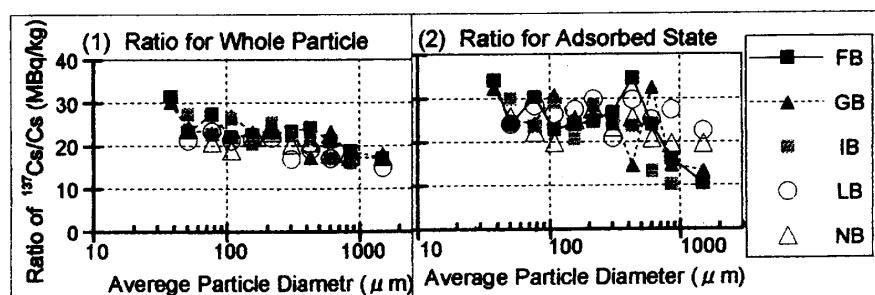


Fig. 9 Ratio of ^{137}Cs to Stable Cs Concentration in the Whole Particle and the Adsorbed Layer

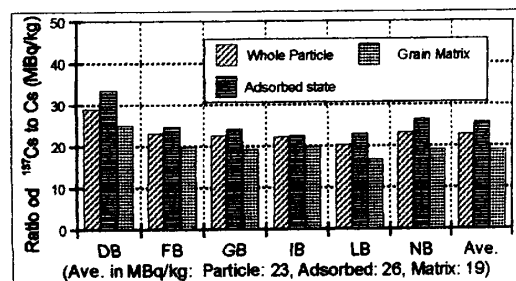


Fig. 10 Comparison of the Concentration of Stable Cs and ^{137}Cs in the Whole Particle and the Adsorbed Layer

and the isotopic exchange between each cesium has been proceeding from the water to the adsorbed layer and the matrix.

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